

REVIEW

A fundamental dichotomy in long-chain polyunsaturated fatty acid abundance between and within marine and terrestrial ecosystems

Stefanie M. Colombo, Alexander Wacker, Christopher C. Parrish, Martin J. Kainz, and Michael T. Arts

Abstract: Polyunsaturated fatty acids (PUFA), especially long-chain (i.e., ≥20 carbons) polyunsaturated fatty acids (LC-PUFA), are fundamental to the health and survival of marine and terrestrial organisms. Therefore, it is imperative that we gain a better understanding of their origin, abundance, and transfer between and within these ecosystems. We evaluated the natural variation in PUFA distribution and abundance that exists between and within these ecosystems by amassing and analyzing, using multivariate and analysis of variance (ANOVA) methods, >3000 fatty acid (FA) profiles from marine and terrestrial organisms. There was a clear dichotomy in LC-PUFA abundance between organisms in marine and terrestrial ecosystems, mainly driven by the C₁₈ PUFA in terrestrial organisms and omega-3 (n-3) LC-PUFA in marine organisms. The PUFA content of an organism depended on both its biome (marine vs terrestrial) and taxonomic group. Within the marine biome, the PUFA content varied among taxonomic groups. PUFA content of marine organisms was dependent on both geographic zone (i.e., latitude, and thus broadly related to temperature) and trophic level (a function of diet). The contents of n-3 LC-PUFA were higher in polar and temperate marine organisms than those from the tropics. Therefore, we conclude that, on a per capita basis, high latitude marine organisms provide a disproportionately large global share of these essential nutrients to consumers, including terrestrial predators. Our analysis also hints at how climate change, and other anthropogenic stressors, might act to negatively impact the global distribution and abundance of n-3 LC-PUFA within marine ecosystems and on the terrestrial consumers that depend on these subsidies.

Key words: climate change, food webs, omega-3, polyunsaturated fatty acids, trophic ecology.

Résumé: Les acides gras polyinsaturés (AGP), particulièrement les acides gras polyinsaturés à longue chaîne (AGPLC), c.-à-d. ≥20 carbones, sont fondamentaux à la santé et à la survie d'organismes marins et terrestres. Donc, il est impératif que nous obtenions une meilleure compréhension de leur origine, de leur abondance et du transfert entre et au sein de ces écosystèmes. Nous avons évalué la variation naturelle de la distribution et de l'abondance des AGP qui existe entre et au sein de ces écosystèmes en amassant et en analysant, par des méthodes multivariables et d'analyse des variations, >3000 profils d'acides gras (AG) d'organismes marins et terrestres. Il y avait une dichotomie claire au niveau de l'abondance des AGPLC entre les organismes dans des écosystèmes marins et terrestres, découlant principalement des AGP C₁₈ chez les organismes terrestres et de l'oméga 3 des AGPLC (n-3) chez les organismes marins. La teneur en AGP d'un organisme dépendait à la fois de son biome (océanique par rapport à terrestre) et de son groupe taxonomique. Au niveau du biome océanique, la teneur en AGP a varié parmi les groupes taxonomiques. La teneur en AGP d'organismes marins dépendait à la fois de la zone géographique (c.-à-d., la latitude; et ainsi largement lié à la température) et du niveau trophique (fonction du régime alimentaire). Les teneurs en AGPLC n-3 étaient plus hautes chez les organismes marins polaires et tempérés que chez ceux des tropiques. Donc, nous concluons que, au prorata de la population, les organismes marins de latitude élevée fournissent une part mondiale importante et disproportionnée de ces substances nutritives essentielles aux consommateurs, y compris les prédateurs terrestres. Notre analyse suggère aussi que le changement climatique, et d'autres agents stressants anthropiques pourraient agir de façon à avoir un impact négatif sur la distribution et l'abondance mondiales d'AGPLC n-3 des écosystèmes marins et sur les consommateurs terrestres qui dépendent de ces contributions. [Traduit par la Rédaction]

Mots-clés: changement climatique, réseaux trophiques, oméga-3, acides gras polyinsaturés, écologie trophique.

Introduction

Fatty acids (FA) are the building blocks of structurally and functionally important lipid molecules (e.g., triacylglycerols, phospholipids) found in all organisms. Their production and distribution varies among plants and animals and among biomes. A subset of FA, the long-chain (i.e., ≥20 carbons long) polyunsaturated fatty acids (LC-PUFA), are known to have key physiological functions in all vertebrate organisms, specifically involved in supporting neurological development

and function, cardiovascular health, visual acuity, growth, reproduction, and the immune system (Brenna et al. 2009; Simopoulos 2011; Swanson et al. 2012; Calder 2015). The LC-PUFA are also critically involved in maintaining structure and fluidity in cell membranes (Arts and Kohler 2009). In particular, eicosapentaenoic acid (EPA; 20:5n–3), docosahexaenoic acid (DHA; 22:6n–3), and arachidonic acid (ARA; 20:4n–6) have distinct and vital functions in vertebrates. EPA has anti-inflammatory effects, lowers the risk

Received 30 June 2016. Accepted 26 October 2016.

S.M. Colombo and M.T. Arts. Department of Chemistry and Biology, Ryerson University, 350 Victoria Street, Toronto, ON, Canada.

A. Wacker. Institute of Biochemistry and Biology, University of Potsdam, Am Neuen Palais 10, Potsdam, 14469 Germany.

C.C. Parrish. Department of Ocean Sciences, Memorial University of Newfoundland, Marine Lab Road, St. John's, NL, Canada.

M.J. Kainz. WasserCluster Lunz, Inter-university Centre for Aquatic Ecosystem Research, 3293 Lunz am See, Austria.

Corresponding author: Stefanie M. Colombo (email: scolombo@ryerson.ca).

Copyright remains with the author(s) or their institution(s). Permission for reuse (free in most cases) can be obtained from RightsLink.

of cardiovascular disease, positively influences immune functions and defense against infections, and may protect against some cancers (reviewed by Calder 2015). DHA is directly involved in several processes in the brain including neurotransmission, cell survival, and prevents neuro-inflammation (and thereby mood and cognition, Bazinet and Layé 2014), and also acts as a precursor for docosatrienes, involved in anti-inflammation (Hong et al. 2003). ARA is also crucial for brain functioning, cell signaling, and is a precursor for endocannabinoids (Turcotte et al. 2015) and eicosanoids (Calder 2015). Thus, together, these three LC-PUFA form an important foundation for health and, by extension, survival in vertebrates.

There are large differences in the LC-PUFA composition of terrestrial and aquatic primary producers (Hixson et al. 2015; Twining et al. 2016). This inherent difference in LC-PUFA production at the base of aquatic and terrestrial food webs has important physiological consequences for consumers. The LC-PUFA are mostly synthesized by primary producers at the base of aquatic food webs (e.g., diatoms, dinoflagellates, cryptophytes; Brett and Müller-Navarra 1997; Taipale et al. 2013; Galloway and Winder 2015). They are progressively consumed, and generally selectively retained, by other aquatic organisms higher up in the food chain (e.g., zooplankton, benthic invertebrates, molluscs, and fish; Dalsgaard et al. 2003; Iverson et al. 2004; Kainz et al. 2004; Budge et al. 2006; Hixson et al. 2015). Even small animals subjected to strong selection pressures (e.g., fasting), and that have limited lipid storage capacity, tenaciously retain these important compounds (Kainz et al. 2004; Schlechtriem et al. 2006; Twining et al. 2016). While vertebrates can synthesize their own LC-PUFA from their dietary precursors (ALA and LNA), the rate of synthesis is generally limited (Cook and McMaster 2004; Tocher et al. 2006); therefore, consuming EPA, DHA, and ARA pre-formed in the diet is highly advantageous for many vertebrates (Arts et al. 2001; Parrish 2009).

Freshwater ecosystems are known to be rich in n-3 PUFA, particularly EPA and DHA, in comparison with terrestrial ecosystems (Hixson et al. 2015; Twining et al. 2016). While the abundance and distribution of marine-derived LC-PUFA has been the focus of a plethora of studies for decades (e.g., Brockerhoff et al. 1963; Ackman et al. 1968; Dalsgaard et al. 2003; Budge et al. 2006; Arts et al. 2009; Parrish et al. 2015), a more global inventory of marine organism PUFA contents has yet to be compiled.

Aquatic organisms require cell membrane FA of varying chain lengths and numbers of double bonds to adapt to ambient water temperatures (Sinensky 1974; Cossins and Prosser 1978; Arts and Kohler 2009). This raises the question, is latitude (as a proxy for temperature) associated with distinct PUFA patterns in marine organisms and, if so, are there climate change implications? This distinction in PUFA content among marine latitudinal zones has been shown in certain taxa, such as phytoplankton (Hixson and Arts 2016), macrophytes (van Ginneken et al. 2011) and zooplankton (Kattner and Hagen 2009), however it has not been documented considering the marine biome as a whole.

While it has been established that there is a distinct functional dichotomy in the LC-PUFA production and abundance between freshwater and terrestrial ecosystems (Hixson et al. 2015), there is also a need to quantitatively establish this distinction between marine and terrestrial ecosystems. The n-3 and n-6 PUFA (ALA, EPA, DHA, and LNA and ARA, respectively) are of special interest due to their essentiality and physiological functions in organisms (Parrish 2009); therefore, we focused on these PUFA. The primary objective of this study is to quantify differences in PUFA among marine and terrestrial organisms, at varying trophic levels, to more rigorously document and quantify the distinct and natural variation in PUFA distribution and abundance that exists between and within these ecosystems.

Methodological approach

Data collection

Fatty acid data from marine and terrestrial organisms were collected from the primary, peer-reviewed, scientific literature. Articles were located and retrieved using the following citation indexing services: Google Scholar©, Web of Science©, and Scholars Portal©. The following search terms were used in combination with "fatty acids": marine, terrestrial, phytoplankton, macrophytes, zooplankton, bivalves, gastropods, annelids, echinoderms, crustaceans, cephalopods, fish, marine mammals, plants, insects, terrestrial mammals. In addition to the data derived from these searches, we also used the marine phytoplankton FA profiles from Hixson and Arts (2016) and bird FA profiles from Galván et al. (2015).

To qualify for inclusion in the data set, the data were required to be species-specific, sampled from either muscle tissue (the greatest tissue mass in the body) or the whole body, and the animals were not cultured (with the exception of cultured phytoplankton and macrophytes). Each study must have presented all FA of interest: ALA, LNA, EPA, DHA, ARA as well as the sums of saturated FA (SFA), monounsaturated FA (MUFA), and PUFA (or total FA along with a complete list of FA to calculate these sums). The FA data must also have been presented as proportional data, i.e., individual FA as a percentage of total FA from total lipids (or the calculated mean of neutral and polar lipid), referred to as "FA content" hereafter. Although it would have been preferable to perform a data synthesis on FA contents expressed as mass-fractions (mg·FA·g-1 wet or dry weight tissue extracted), the majority of studies (>75%) present FA data on a proportional basis (i.e., percentage). A Grubb's outlier test was used to determine if there were significant outliers (p < 0.05) for each FA within a particular group. If a significant outlier was detected (which occurred 8 times in the entire data set), the data entry was removed from the dataset.

Dataset

The FA data, for marine and terrestrial organisms, were sorted into 14 major taxonomic groups. The marine organisms comprised the following 11 taxonomic groups: phytoplankton, macrophytes, zooplankton, bivalves, gastropods, annelids, echinoderms, crustaceans, cephalopods, fish, birds, and marine mammals. The terrestrial taxa were sorted into 4 major taxonomic groups: plants, insects, birds, and terrestrial mammals. Complete FA data on other terrestrial vertebrates, such as reptiles and amphibians, are scarce in the literature and thus were not included. To investigate patterns within either marine or terrestrial food webs, the abovementioned groups were pooled, as follows, based on their approximate trophic position. Primary producers consisted of phytoplankton, macrophytes, and terrestrial plants. Herbivores/omnivores consisted of annelids, bivalves, herbivorous/omnivorous crustaceans, zooplankton, and herbivorous/omnivorous terrestrial mammals. Carnivores consisted of cephalopods, carnivorous crustaceans, carnivorous fish, marine mammals, and carnivorous terrestrial mammals. The marine FA data were also sorted by geographic origin. This was determined by the location where the organism was sampled (as listed in the original primary article) and which was characterized by three latitudinally-defined geographic zones: polar (90-60 °N and °S), temperate (60-30 °N and °S), and tropical (30 °N to 30 °S). In cases where algal cultures were obtained from a collection, we used Algaebase (http://www.algaebase.org/) and (or) the World Register of Marine Species (WoRMS, http://www. marinespecies.org/index.php) to obtain information on the known global distribution (range) of that algal species. Algae were considered "cosmopolitan" when they were listed as occupying two or more, of our three, latitudinal zones.

Analysis of variance

ANOVA models were used to determine whether biome, taxonomic group, latitudinal zone, and (or) trophic level could ac-

Table 1. ANOVA models used to analyze fatty acid content (% total fatty acids) in marine and terrestrial organisms.

Model number	Model type	Variables	Objective
1	ANOVA	Biome (categorical) Taxonomic Group (categorical) Biome × Group Interaction	To determine the effect of biome, taxonomic group, and their interaction on fatty acid content in all organisms
2	ANOVA	Biome (categorical) Trophic level (categorical)	To determine if fatty acid content in organisms depends on increasing trophic level in either biome
3	ANOVA	Taxonomic group (categorical) Latitudinal zone (categorical)	To determine the effect of taxonomic group and latitudinal zone on the fatty acid content of marine organisms
4	ANOVA	Latitudinal zone (categorical) Trophic level (categorical)	To determine if fatty acid content in marine organisms depends on increasing trophic level among latitudinal zones
5	ANOVA (nested)	Trophic level (categorical) Taxonomic group (nested factor)	To determine whether trophic level or taxonomic group (within each trophic level) are more important in predicting FA content in marine organisms ^a
6	ANOVA (nested)	Taxonomic group (categorical) Latitudinal zone (nested factor)	To determine whether the fatty acid content in marine organisms differs depending on latitudinal zone within each taxonomic group

"In this model we were not interested in the distinct taxonomic group, but rather the difference among random groups within each trophic level.

Table 2. Fatty acid content (% total fatty acids) of marine and terrestrial organisms (mean ± SE).

Taxonomic group	LNA	ALA	ARA	EPA	DHA	Σ SFA	Σ MUFA	ΣPUFA	Σn-3 FA	Σn-6 FA	DHA/EPA
Marine (n = 2063)											
Phytoplankton ($n = 504$)	2.5±0.2	3.5±1.0	0.8 ± 0.1	11.9±0.4	5.2±0.3	37.0±0.7	25.3±0.5	35.4±0.7	20.8±1.1	3.3 ± 0.2	3.6±0.5
Macrophytes ($n = 191$)	6.2±0.4	6.6±0.5	8.6±0.5	13.2±1.0	0.2 ± 0.1	35.0±1.0	19.8±0.7	44.7±1.1	20.1±0.1	14.8±0.6	0.05±0.01
Zooplankton ($n = 139$)	1.7±0.1	1.6±0.1	1.2±0.2	16.2±0.5	15.2±0.7	24.9±0.8	32.1±1.5	44.6±1.2	33.0±1.1	2.9±0.2	1.0±0.05
Bivalves $(n = 242)$	2.1±0.1	1.8±0.1	2.4 ± 0.1	14.6±0.5	11.9±0.5	30.7±0.9	20.5±0.6	46.2±0.9	28.3±0.8	4.4±0.2	1.1±0.1
Gastropods ($n = 118$)	2.6±0.2	2.4±0.2	7.8±0.5	10.0±0.8	2.5±0.5	30.9±0.8	23.5±0.6	41.6±1.0	14.9±1.0	10.4±0.5	0.8 ± 0.2
Annelids $(n = 48)$	5.2±1.0	1.9±0.3	2.2±0.2	9.7±1.0	3.3±0.6	24.8±0.9	32.8±1.6	37.6±1.8	14.9±1.3	7.4±1.0	0.3±0.1
Echinoderms ($n = 106$)	2.6±1.2	1.4±0.2	11.2±0.7	13.0±0.8	3.8 ± 0.4	26.2±1.1	29.2±0.9	41.4±1.5	18.2±1.0	13.8±1.4	0.8 ± 0.4
Crustaceans ($n = 153$)	1.9±0.2	0.6±0.1	4.6±0.3	15.3±0.6	10.7±0.5	24.2±0.8	28.7±0.9	39.2±1.1	26.7±0.9	6.5±0.4	0.8 ± 0.04
Cephalopods $(n = 118)$	0.6±0.1	0.3 ± 0.0	3.3±0.3	12.8±0.4	23.9±0.9	28.2±0.9	16.3±1.1	44.7±1.4	37.0±1.2	4.0±0.3	1.9±0.1
Fish $(n = 239)$	1.5±0.1	0.9±0.1	2.2 ± 0.2	7.7±0.3	17.7±0.6	29.2±0.5	34.0±0.8	34.9±0.8	26.3±0.7	3.7±0.2	2.9±0.1
Mammals $(n = 205)$	1.6±0.1	0.7 ± 0.1	1.0±0.3	5.4±0.2	7.9±0.3	18.1±0.4	55.6±0.7	27.5±1.1	14.1±0.4	2.6±0.3	1.7±0.1
Terrestrial $(n = 1009)$	Terrestrial $(n = 1009)$										
Plants $(n = 410)$	38.4±1.0	11.8±0.8	0.5 ± 0.1	_		20.9±0.8	26.6±0.9	50.5±1.0	11.8±0.8	38.9±1.0	_
Insects $(n = 389)$	15.4±0.7	13.4±0.8	16.5±2.9	1.3±0.5	_	35.4±1.0	35.8±0.9	28.8±1.0	14.7±0.1	31.9±2.6	
Birds $(n = 107)$	9.96±0.5	0.99±0.1	8.18±0.3	0.14 ± 0.0	3.75±0.2	48.4±0.5	25.6±0.7	26.0±0.6	6.3±0.2	19.8±0.6	37.5±2.1
Mammals $(n = 193)$	12.5±0.8	5.6±1.2	2.6±0.5	0.9±0.1	1.8±0.3	40.3±1.1	36.8±1.1	22.5±1.2	8.3±0.6	15.1±0.7	2.1±0.4

count for the variance in FA content among the organisms in our dataset. Biome (a fixed categorical variable) included 2 groups (marine or terrestrial). Taxonomic group (also a fixed categorical variable) included 11 groups in the marine biome (see above) and 3 groups in the terrestrial biome (see above). Latitudinal zone (a random categorical variable) was characterized by three geographic regions (see above). Trophic level (a random categorical variable) had three levels: primary producers, herbivore/omnivores, and carnivores, as above. A nested ANOVA was used to determine the effect of latitudinal zone within taxonomic group.

Using a combination of the above factors, 6 statistical models (using Minitab Statistical Software, version 17) were tested on each individual PUFA or FA group (e.g., SFA, MUFA, PUFA) to determine the effects and (or) interactions among biome, taxonomic group, latitudinal zone, and trophic position on FA content (Table 1). The categorical factors were considered to have significant effects on FA proportions and (or) distributions when p < 0.05, and, where significant differences occurred, treatment means were further differentiated using the Tukey HSD multiple comparison test, with Bonferroni corrections. The residuals from each model were examined to ensure there were no violations of model assumptions. Plots of residuals were evaluated for model assumptions: homogeneity of variance (plot of residuals vs. fitted values), normality (frequency of residuals, normal probability plot), and independence assumptions (residuals vs. lag + 1 residuals). Levene's test was also used to evaluate equal variances.

Multivariate analyses

All multivariate analyses were conducted using PRIMER-E (Plymouth Routines in Multivariate Ecological Research; PRIMER-E Ltd, version 7.0.10, Aukland, NZ). Principal coordinates analysis (PCO) was used to visualize and quantify patterns observed among FA profiles (individual PUFA and the sums of SFA, MUFA, and PUFA) of organisms depending on their biome (marine or terrestrial), taxonomic group (see above), and approximate trophic level (see above). To evaluate the significance of the apparent biome, taxonomic group, and trophic level separation observed in the PCO, we used a one-way permutational multivariate ANOVA (PERMANOVA) on the same FA dataset. The PCO was used on the entire data set of PUFA and FA groups (in both marine and terrestrial organisms), as well as the marine data subset, the terrestrial data subset, and the phytoplankton and fish data subsets. The FA vectors included all those in the analysis, listed above.

The Similarity Percentage analysis (SIMPER) was used to differentiate FA profiles based on biome and taxonomic group. The similarity coefficient in SIMPER ranges from 0% to 100% with the ends of the range representing the extreme possibilities, i.e., S = 0% if two samples are totally dissimilar and S = 100% if two samples are totally similar. The non-metric Bray–Curtis dissimilarity statistic was used to quantify the compositional dissimilarity between samples in the PCO and SIMPER (Bray and Curtis 1957). This test delivers robust and reliable dissimilarity results and is one of the most commonly used metrics to explore relationships in ecology, environmental sciences, and related fields (Clarke and Warwick

Table 3. Fatty acid content (% total fatty acids) of marine and terrestrial primary producers, herbivores/omnivores, and carnivores (mean ± SE).

	,		,			<i>D</i> 1		,		,	,
Trophic level	ALA	LNA	ARA	EPA	DHA	Σ SFA	Σ MUFA	ΣPUFA	Σ n-3 FA	Σn-6 FA	DHA/EPA
Marine											
Primary producers ¹	$4.83\pm0.8^{\rm a}$	3.59±0.2a	3.17 ± 0.2^{b}	12.3±0.4 ^b	3.84 ± 0.2^{c}	36.4±0.6a	23.7±0.4°	38.1±0.6 ^b	23.7±0.4 ^b	6.85±0.3a	2.5±0.3a
Herbivores/omnivores ²	$1.61\pm0.1^{\rm b}$	2.34 ± 0.2^{b}	4.54 ± 0.2^{a}	13.8±0.3a	9.18±0.3 ^b	27.5±0.4 ^b	26.3±0.4 ^b	42.9±0.5a	26.3±0.5 ^b	6.88±0.3a	$0.9\pm0.1^{\rm b}$
Carnivores ³	$0.67\pm0.0^{\rm b}$	1.37±0.1 ^c	2.09±0.2 ^c	8.18±0.2c	15.4±0.4a	25.1±0.4°	37.8±0.6a	34.4±0.6c	38.1±0.6a	3.45 ± 0.2^{b}	2.2±0.1a
Terrestrial											
Primary producers ¹	11.1±0.8a	38.4±1.0a	0.52±0.1 ^c	0.00 ± 0.0^{c}	$0.00\pm0.0^{\rm b}$	20.8±0.8c	52.5±2.1a	50.3±1.0a	26.6±0.9c	36.4±1.5a	
Herbivores/omnivores ²	8.72±0.5 ^b	14.1±0.5 ^b	8.07±0.8a	0.62 ± 0.1^{b}	3.12±0.2a	39.2±0.7a	33.4 ± 0.7^{b}	27.5±0.7 ^b	33.4 ± 0.7^{b}	$18.4 \pm 0.7^{\rm b}$	5.0 ± 0.2^{a}
Carnivores ³	6.92±3.6°	7.87±0.9°	2.35 ± 0.7^{b}	2.39±0.4a	3.25±0.4 ^a	25.4±1.6 ^b	26.6±0.9°	20.6±1.2°	52.5±0.2a	10.2±1.4°	1.6 ± 0.2^{b}
Biome term (p-value)	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	0.819	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001
Trophic term (p-value)	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	0.009	< 0.001	< 0.001

Note: Different superscripts in the same column (i.e. for each fatty acid) indicates significant differences among trophic levels within each biome.

1Primary producers: phytoplankton, macrophytes, and terrestrial plants.

Table 4. Fatty acid content (% total fatty acids) of marine organisms in three latitudinal zones: polar (90–60 °N and °S), temperate (60–30 °N and °S), and tropical (30 °N to 30 °S).

Latitudinal zone	ALA	LNA	ARA	EPA	DHA	Σ SFA	Σ MUFA	Σ PUFA	Σ n-3 FA	Σn-6 FA	DHA/EPA
Polar $(n = 372)$	1.7±0.2	2.1±0.2	1.9±0.3a	13.6±0.5a	10.5±0.4a	22.2±0.5a	39.9±0.9a	42.8±0.8a	24.5±0.7a	4.1±0.3a	1.9±0.2
Temperate ($n = 1217$)	2.6 ± 0.4	2.6 ± 0.1	3.8 ± 0.1^{b}	12.3±0.2 ^b	9.7 ± 0.3^{a}	30.1 ± 0.4^{b}	26.3±0.4 ^b	40.1±0.4 ^b	24.6±0.6a	6.4 ± 0.2^{b}	1.5±0.1
Tropical $(n = 263)$	2.3 ± 0.3	3.1 ± 0.2	4.6 ± 0.3^{c}	6.9 ± 0.4^{c}	8.3±0.6 ^b	35.1±0.8c	26.3±0.7 ^b	34.7±0.8°	18.1±0.7 ^b	$7.7\pm0.4^{\circ}$	2.5±0.3

Note: Different superscripts in the same column (i.e. for each fatty acid) indicate significant differences among latitudinal zones.

Table 5. Fatty acid content (% total fatty acid; mean \pm SE) in marine and terrestrial organisms, with results of ANOVA (model 2; biome and trophic level as categorical variables).

			Biome		Trophi	c level	
Fatty acid	Marine	Terrestrial	F-stat	<i>p</i> -value	F-stat	<i>p</i> -value	
ALA	2.3±0.3	9.6±0.5	177	<0.001	20.1	< 0.001	
LNA	2.5±0.1	23.6±0.6	2121	< 0.001	272	< 0.001	
ARA	3.4±0.1	7.2±0.6	52.2	< 0.001	37.1	< 0.001	
EPA	11.7±0.1	0.9±0.1	405	< 0.001	71.5	< 0.001	
DHA	9.2±0.2	2.1±0.1	167	< 0.001	329	< 0.001	
Σ SFA	29.8±0.3	31.6±0.6	0.05	0.819	48.6	< 0.001	
Σ MUFA	28.8±0.3	31.8±0.6	101	< 0.001	210	< 0.001	
Σ PUFA	38.9±0.3	35.8±0.7	43.3	< 0.001	86.5	< 0.001	
Σ n-3 FA a	23.2±0.4	6.3±0.3	202	< 0.001	12.3	0.009	
Σ n-6 FA ^b	5.7±0.2	17.6±0.7	482	< 0.001	53.2	< 0.001	

aSum of n-3 = ALA + EPA + DHA.

2001). A permutational analysis of homogeneity of dispersions (PERMDISP) was used to test the homogeneity of multivariate dispersions within groups. Significant differences in dispersion within taxonomic groups were observed (p = 0.001), therefore data were square-root transformed prior to multivariate analysis (Greenacre and Primicerio 2013; Hixson et al. 2015).

Data synthesis results

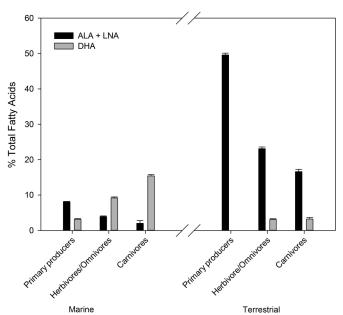
Database

A total of 3072 FA profiles were collected from marine and terrestrial organisms, which were subsequently divided into biomes (Table 2), taxonomic groups (Table 3), and latitudinal zones (in the case of marine organisms only; Table 4). The data set used in the analyses is provided in the Supplementary data¹.

ANOVA results

Both biome and taxonomic group, and the interaction between them, significantly determined the FA content of organisms (Supplementary data, Table S1)¹. The single exception was that the taxonomic group did not determine SFA; however, because the

Fig. 1. Sum of ALA \pm LNA and DHA (mean \pm SE) by trophic level in marine and terrestrial organisms.



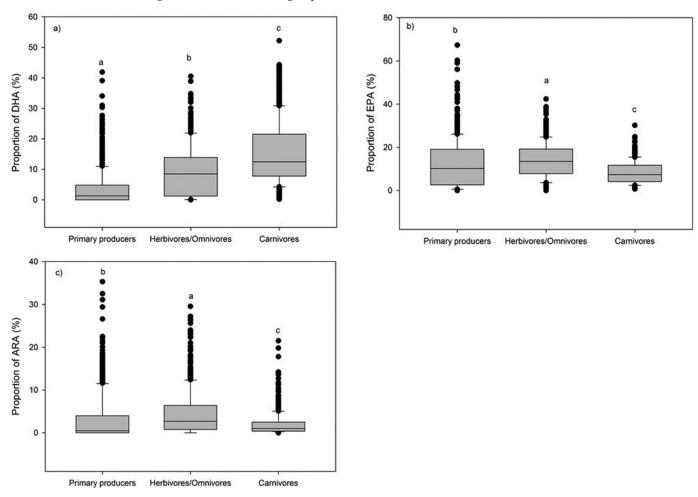
interaction between biome and taxonomic group was significant for SFA, taxonomic group could not be interpreted alone (Supplementary data, Table S1)¹. Trophic level determined PUFA and FA groups, as values tended to increase or decrease with increasing trophic level in both biomes (Table 5). This pattern was most striking in ALA, LNA, and DHA when compared between marine and terrestrial organisms (Fig. 1). ALA plus LNA decreased with increasing trophic level, but the magnitude of the change was different in marine and terrestrial ecosystems. DHA was not observed in terrestrial plants, and was also not recorded in insects (Table 2, Fig. 1). DHA content generally increased from primary producers to carnivores in marine ecosystems, with carnivores

²Herbivores/omnivores: annelids, bivalves, certain crustaceans, echinoderms, zooplankton, terrestrial insects, and certain terrestrial mammals.

³ Carnivores: cephalopods, certain crustaceans, fish, marine mammals, and certain terrestrial mammals.

 $^{^{}b}$ Sum of n-6 = LNA + ARA.

Fig. 2. Boxplots of (a) DHA, (b) EPA, and (c) ARA levels in marine organisms by trophic level (primary producers to carnivores), where p < 0.001 and different letters indicate significant differences among trophic levels for each LC-PUFA.



having significantly higher proportions than the other groups (Fig. 2a). ARA was higher in terrestrial organisms compared to marine organisms (Table 5). Within the marine ecosystem, herbivores/omnivores had higher EPA (Fig. 2b) and ARA contents (Fig. 2c) than primary producers and carnivores. ARA contents varied greatly among taxonomic groups within either ecosystem, and were notably high in macrophytes, gastropods, echinoderms, and insects (Table 2).

Latitudinal zone explained the variation in PUFA content for all PUFA and FA summary categories in marine organisms, except ALA and LNA (Supplementary data, Table S2)¹. Taxonomic group within marine environments also determined FA content (two-way ANOVA). Within each taxonomic group, latitudinal zone remained a significant factor in FA content of marine organisms, regardless of grouping (nested ANOVA; Supplementary data, Table S3)¹. In addition to latitudinal zone (except for ALA and LNA), individual PUFA content also depended on trophic level, and generally increased with higher trophic levels (Supplementary data, Tables S4, S5¹; Fig. 3). A nested ANOVA revealed that taxonomic group within each trophic level determines FA content (p < 0.05 for all FA, except ALA; Supplementary data, Table S6)¹, not just trophic level alone (p > 0.05 for all FA).

Multivariate analyses results

There was a clear separation, using PCO, between marine vs. terrestrial organisms along PCO1 (Fig. 4), confirming that biome was a significant factor in the similarity matrix, according to the PERMANOVA design (F = 1110.6; p = 0.001; number of permutations = 999). The ordination explains 62.4% of the variation; PCO1

explained most of the total variation (40.3%), and PCO2 explained another 22.1%. The FA vectors along PCO1 indicated an association among EPA, DHA, and ARA with marine organisms, and ALA and LNA with terrestrial organisms. Along PCO2, SFA and MUFA showed an opposing relationship to PUFA (this was expected as these sums add up to 100%) but did not divide the plot according to the biome factor imposed in this particular plot.

PCO explained 51.8% of the total variation (PCO1 = 30%, PCO2 = 21.8%) in the FA content of marine organisms as a function of latitudinal zone. Taxonomic group (F = 6.31; p < 0.001), latitudinal zone (F = 18.8; p < 0.001), and the interaction between taxonomic group and latitudinal zone (F = 5.03; p < 0.001) were significant factors in the Bray-Curtis similarity matrix according to PERMANOVA (with 999 unique permutations). Using marine phytoplankton as a subset, the ordination explained 80.8% of the total variation, with PCO1 explaining most of the variation (52.4%), and PCO2 the remainder (28.4%). The PERMANOVA analysis revealed that latitudinal zone (p = 0.002), taxonomic group (p = 0.001), and the interaction between zone and group (p = 0.001) were significant factors in the Bray-Curtis similarity matrix. For marine fish (Fig. 5), the ordination explained 74.3% of the total variation (PCO1 explained 55.7%, and PCO2 explained 18.6% of the variation). Latitudinal zone was a significant factor in the Bray-Curtis similarity matrix according to PERMANOVA (F = 16.0; p = 0.001; with 999 unique permutations). Within the terrestrial organism data set, PCO1 explained 40.5% of total variation, and PCO2 explained 19.5% (Fig. 6). Taxonomic group

Fig. 3. Three-dimensional representation of DHA (%) in marine organisms organized by trophic level (x-axis) and latitudinal zone (z-axis). See Supplementary data Table S5 for mean values (±SE) and Table S6 ANOVA model p-values.¹ Different letters indicate significant differences among each factor, trophic level (applies to DHA in three trophic levels, regardless of latitudinal zone) and latitudinal zone (applies to DHA in three zones, regardless of trophic level). [Colour online.]

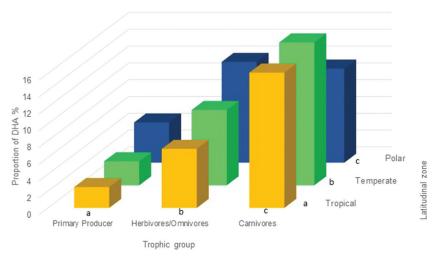


Fig. 4. Principal coordinates analysis plot of fatty acids in marine and terrestrial organisms (n = 3072), using a Bray–Curtis similarity matrix (square-root transformed). [Colour online.]

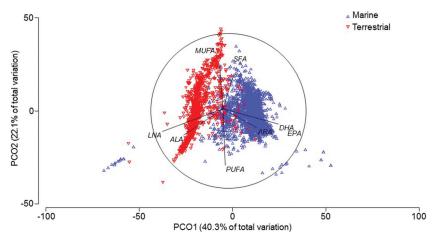


Fig. 5. Principal coordinates analysis of fatty acids in marine fish (n = 239) by latitudinal zone, using a Bray–Curtis similarity matrix (square-root transformed). [Colour online.]

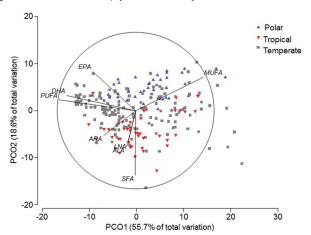
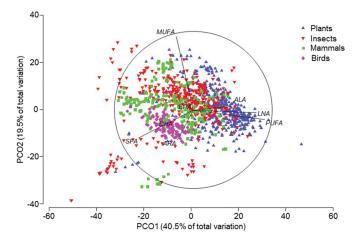


Fig. 6. Principal coordinates analysis of fatty acids in terrestrial organisms (n = 1009) by taxonomic group, using a Bray–Curtis similarity matrix (square-root transformed). [Colour online.]



was a significant factor in the Bray–Curtis similarity matrix according to PERMANOVA (F=113.2; p=0.001; with 995 unique permutations). The FA vectors along PCO1, DHA, and ARA, indicated an association with birds and mammals, as opposed to plants and insects. Conversely, ALA, LNA, and PUFA were associated with plants and insects along PCO1.

SIMPER analysis revealed that within the marine data set, macrophytes and marine mammals were the most dissimilar pair based on their FA content (30.8% dissimilar), while fish and zooplankton were the most similar pair (16.1%). In both comparisons, MUFA caused the most variation between groups. Within taxonomic groups, phytoplankton were the most diverse in terms of FA content (71.8% similarity within the group) and marine mammals were the most similar (88.2% similarity within the group). Dispersion among groups was not homogenous (p = 0.001) according to PERMDISP, which is partially explained by unequal sample sizes among taxonomic groups, with some groups (such as phytoplankton) with a large number of samples (n = 505) and large dispersion and higher variation. Within the marine data set, latitudinal zones were equally dissimilar (22.6%-23.7%). DHA caused the greatest dissimilarity between temperate and tropical marine organisms, while MUFA caused the most dissimilarity between polar and tropical or temperate marine organisms. Within the terrestrial data set, plants and mammals were the most dissimilar (34.5%), followed by plants and birds (33.2%), insects and mammals (30.6%), plants and insects (29.3%), insects and birds (29.1%), and birds and mammals (25.8%). LNA and ARA caused the most dissimilarities among all terrestrial groups compared.

Summary and perspective

Data synthesis

Our data synthesis, consisting of 3072 FA profiles from marine and terrestrial organisms, allowed us to uncover a striking dichotomy in the distribution and abundance of PUFA between marine and terrestrial ecosystems. This contrast was primarily driven by the n-3 LC-PUFA (EPA, DHA) content in marine organisms, compared with terrestrial organisms (which contained higher levels of the C₁₈ PUFA ALA and LNA; Fig. 4). Marine organisms also had higher total n-3 PUFA contents, while terrestrial organisms had typically high total n-6 PUFA contents. While the FA content of marine organisms has been documented for hundreds of individual species, and the difference between freshwater and terrestrial organisms has been established (Hixson et al. 2015), this is the first time FA composition in marine and terrestrial ecosystems has been compared in a comprehensive quantitative study.

Our data set represented 14 taxonomic groups distributed among marine and terrestrial ecosystems. While our data collection was as thorough as possible, FA profiles for marine organisms were much more prevalent in the literature than those for terrestrial organisms, and within marine organisms certain taxa were better represented than others (often based on their relevance to human consumption; e.g., bivalves and fish). Grouping species together (i.e., by taxonomic group and (or) trophic level) also introduced a degree of variability, as phylogeny is a known driver of FA content (Budge et al. 2002; Dalsgaard et al. 2003; Makhutova et al. 2011; Galloway and Winder 2015; Hixson et al. 2015). While we organized our data taxonomically and by trophic level to aid in the detection and interpretation of FA patterns among them, there is still marked variation in a single species within a taxonomic group because of an organism's specific habitat, feeding preferences, gender, life stage, etc., even when these species shared the same biome and had similar diets (according to trophic position). This introduced variation into our analysis. However, despite species differences, the variation within groups was minimal compared to the differences between groups, particularly between biomes. Our main conclusion is that a fundamental dichotomy exists in the relative abundance, and therefore global distribution, of PUFA between marine and terrestrial organisms.

Dichotomy between biomes

The difference in FA content between marine and terrestrial organisms is evident when observed visually in multivariate space (Fig. 4). This disparity is mainly driven by the LC-PUFA (EPA, DHA, and ARA) in marine organisms, and the C₁₈ PUFA (ALA and LNA) in terrestrial organisms. This difference in FA content between the two ecosystems has its primary origin at the base of the food webs. In terrestrial ecosystems, ALA was the only n-3 PUFA recorded in primary producers (vascular plants) in our data set (although they are also known to produce other n-3 PUFA, such as 16:3n-3 and 18:4n-3: Dubois et al. 2007). By contrast, in marine ecosystems, primary producers (phytoplankton and macrophytes) predominantly contained n-3 LC-PUFA (EPA and DHA). In addition, our analyses revealed that marine primary producers synthesize, on average, \sim 1.7 x more n-3 PUFA (of the reduced subset of PUFA we analyzed here) than terrestrial primary producers. Our data analysis highlights that marine primary producers consist of 13%-17% dietary n-3 LC-PUFA on average, which is available to subsequent consumers (Table 2), whereas such an n-3 LC-PUFA supply is entirely missing (at least in the plants we surveyed) from terrestrial primary producers. Further, we found that in marine organisms the n-3 PUFA dominate over n-6 PUFA, while the opposite is true for terrestrial organisms, which has also been observed when comparing terrestrial with freshwater organisms (Skjervold 1992; Koussoroplis et al. 2008; Hixson et al. 2015). Within each biome, taxonomic group played an important role, and so did the interaction between an organism's biome and their taxonomic group. Thus, the FA content of a particular organism depends on both an organism's biome (land vs. ocean), and the taxonomic group to which it belongs.

The overall FA composition in organisms has been confirmed experimentally to also depend on their diet (Iverson et al. 1997; Lane et al. 2006; Koussoroplis et al. 2008; Torres-Ruiz et al. 2010) and has been reviewed (Iverson et al. 2004; Budge et al. 2006; Kelly and Scheibling 2012). We found that within each biome, individual PUFA tend to either increase or decrease as a function of an organism's trophic level as well as habitat. For example, the sum of ALA and LNA generally decreases with increasing trophic level (Fig. 1) but are higher in terrestrial food webs than marine. On the other hand, DHA tended to increase with increasing trophic level, at least in the marine biome (Figs. 1, 2, 3). This result was also observed between terrestrial and freshwater organisms (Hixson et al. 2015). Interestingly, while SFA content depended on approximate trophic level, it was not different between biomes.

Similarly to what was observed in our marine data set, other researchers have also observed that DHA was transferred and selectively retained in the aquatic food web and generally increased in higher trophic levels (Kainz et al. 2004; Persson and Vrede 2006; Koussoroplis et al. 2008; Hixson et al. 2015; Twining et al. 2016). The essential PUFA are transferred along trophic levels at about twice the efficiency of bulk carbon (Gladyshev et al. 2011; Hartwich et al. 2013), and they are thereby generally retained, rather than diluted, in the biomass of organisms of higher trophic levels, a characteristic known in freshwater ecosystems (Gladyshev et al. 2013; Hixson et al. 2015; Twining et al. 2016). This is because DHA (and to a lesser extent EPA) is minimally modified from ingestion to assimilation and generally conserved in its original form to serve specific and important physiological purposes, rather than being catabolized and (or) modified (Kainz et al. 2004; Twining et al. 2016). However, it is worth noting that this pattern may be further modified by taxonomic affiliation, as organisms may share a similar trophic position (e.g., cladocerans and copepods) but have markedly different DHA contents. Ultimately, it is taxonomic group within each trophic level that is the most important determinant, as shown in the nested ANOVA model. Conversely, C₁₈ PUFA (ALA and LNA) generally decrease with increasing trophic level (in both aquatic and terrestrial ecosystems), which

suggests that ALA and LNA are utilized either as precursors for the LC-PUFA or catabolized, rather than being retained in tissues.

In terrestrial ecosystems, the EPA and DHA content in carnivores originated either directly from aquatic food sources (see above) or is produced endogenously, although the rate of synthesis from ALA is typically very low in vertebrates (Brenna et al. 2009; Domenichiello et al. 2015). Within terrestrial ecosystems, the FA contents of mammals and birds were different from terrestrial plants, on account of the difference in LNA (higher levels found in plants) and DHA (not found in the plants in our data set; Fig. 6). The ability for vertebrates to either access aquatic food sources or modify ALA becomes important for survival. Therefore, we conclude that terrestrial ecosystems are, in general, a much less important source of n-3 LC-PUFA production than aquatic ecosystems, and that aquatic organisms (both marine and freshwater), through a variety of pathways (Gladyshev et al. 2009), provide an important nutrient subsidy to terrestrial organisms. The availability of rich sources of EPA and DHA is critical for many terrestrial vertebrates, and the timing and quantity may be important as well. For example, shorebirds are known to gorge on n-3 LC-PUFA rich marine resources for long distance migration, i.e., n-3 LC-PUFA are used as "performance enhancing substances" to prepare muscles for migration (Maillet and Weber 2006; Guglielmo 2010).

Fatty acid dynamics in food webs

The LC-PUFA have different physiological functions, may be metabolized and incorporated into tissues differently, and may vary in terms of importance in nutritional requirements for different species (Emery et al. 2016). We found that individual PUFA contents varied as a function of the biome of the organism and (or) their approximate trophic position in the food chain. Generally, we found that FA tended to increase or decrease with increasing trophic level (Fig. 2), although this was not always the case. For example, relative EPA content increased from marine primary producers to herbivores/omnivores, but decreased from herbivores/ omnivores to carnivores. Some of the herbivores/omnivores in our marine data set were zooplankton, which have been shown to exhibit an EPA-retentive metabolism (Kainz et al. 2004; Wacker and Martin-Creuzburg 2007; Hartwich et al. 2013). EPA is highly retained in zooplankton, whereas DHA is highly retained in most fish (e.g., salmonids; Kainz et al. 2004), which explains the different relationships we observed between EPA and DHA in our data. In fish, for example, there is increasing evidence to support the observation that EPA is preferentially catabolized or further converted to DHA (e.g., in salmonids; Murray et al. 2014) rather than spared and retained in tissues, and as such, EPA appears to be generally more dispensable compared to DHA (Trushenski et al. 2012; Glencross et al. 2015; Emery et al. 2016). This was observed in our dataset, which documented a decrease in EPA content in carnivores. It has been suggested that DHA (and ARA) are the primary drivers of LC-PUFA essentiality in some fish species (Trushenski et al. 2012; Rombenso et al. 2016). This may be the case in other aquatic vertebrates and semi-aquatic mammals as well (Koussoroplis et al. 2008), as we observe selective retention in DHA up to carnivores, coupled with a significant drop in EPA content. Therefore, it appears that EPA and DHA are required in different proportions by different consumers and together are not necessarily progressively retained from primary producers up to carnivores, although this is highly dependent on species.

Fatty acid abundance and distribution within marine ecosystems

The diversity of organisms in our data set allowed us to take a quantitative approach in defining similarities and differences in the abundance and distribution of FA within the marine ecosystem. Marine macrophytes and marine mammals had the most dissimilar FA profiles, while fish and zooplankton had the most

similar FA profiles. In both comparisons, MUFA was the underlying source of most of the variation between groups. This distinction is important, as individual MUFA are often recognized as biochemical markers in marine food webs and are frequently used to quantitatively document predator-prey interactions (Iverson 2009). Beyond diet, taxonomy is an important factor that defines FA content in marine organisms (Gladyshev et al. 2013; Galloway and Winder 2015). Species grouped in similar, yet broad taxonomic classifications, did not necessarily possess the same FA content as one another. For example, we found that phytoplankton were the most diverse group in terms of FA content. Dispersion among taxonomic groups was not homogenous, which can be partially explained by unequal sample sizes, with some taxa (such as phytoplankton) that had a large number of samples, showing a large dispersion and higher variation. Ultimately, we found that assuming the trophic level of an organism (as a function of diet, as shown by Iverson et al. 2004; Budge et al. 2006) could be used as a tool to roughly estimate FA content. However, it is taxonomic group within each trophic level that is a more important determinant (Galloway and Winder 2015), as shown in the nested ANOVA model.

Fatty acid differences in marine ecosystems defined by latitude

Temperature directly influences metabolism, functioning, and survival in aquatic organisms (Gaston 2003). In an attempt to maintain physiological homeostasis when faced with temperature changes (Clarke 2003), ectotherms may change the structure of their cell membrane by modifying FA chain length and degree of unsaturation in membrane lipids to maintain a desired level of order (fluidity) in cell membranes (Sinensky 1974; Guschina and Harwood 2006; Arts and Kohler 2009). To maintain fluidity in cold waters, ectotherms tend to retain or increase the content of membrane PUFA, as the double bonds enhance the ability of FA to "bend" and increase membrane fluidity (Hazel 1995; Arts and Kohler 2009; Sperfeld and Wacker 2012; Parrish 2013). Therefore, using latitude (as a proxy for habitat temperature) to quantitatively define the FA content of marine organisms was of interest. We found that geographic zone, taxon, and trophic level synergistically determined the FA content of marine organisms. The C_{18} PUFA (ALA and LNA) content in marine organisms did not differ by geographic zone. These PUFA are essential to animals regardless of latitude, however they differed quantitatively among taxa and

Generally it is assumed that LC-PUFA, especially EPA and DHA, are less prevalent or abundant in tropical marine ecosystems compared to polar marine ecosystems, where EPA and DHA are known to be in rich and abundant supply (Lands 1982; Broadhurst et al. 2002). Tropical organisms tend to possess FA profiles that are somewhat different from those of temperate and polar regions (Ahlgren et al. 2009; Iverson 2009). For example, in polar oceans, marine organisms characteristically contain relatively high levels of MUFA (particularly 20:1 and 22:1), which are often regarded as biomarkers of zooplankton, such as krill and copepods (Brett et al. 2009) These MUFA are routinely transferred to higher trophic level organisms. This was observed in the polar marine organisms in our data set (Table 4), including fish (Fig. 3), relative to tropical and temperate organisms.

We found that the relative proportion of n-3 LC-PUFA, total MUFA, and total PUFA were higher in organisms located in polar regions than those in tropical regions. This aligns with the hypothesis that cold-water ectotherms organize long-chain, unsaturated FA to maintain membrane fluidity to inhabit cold waters (Sinensky 1974; Arts and Kohler 2009). For this reason, having elevated n-3 LC-PUFA levels may be a mechanism used to survive in polar regions, where organisms typically live at temperatures near the freezing point for much of the year (Thomas and Dieckmann 2002). Both in macrophytes (van Ginneken et al. 2011)

and microalgae (Boelen et al. 2013; Hixson and Arts 2016), PUFA content and degree of unsaturation in polar oceans were higher than in tropical oceans. Conversely, total SFA and ARA increase in marine organisms going from polar to tropical regions (assuming increasing temperature); a result also observed in a large data synthesis of phytoplankton (Hixson and Arts 2016) and macrophytes (van Ginneken et al. 2011). It is important to note that we did not consider depth as a factor within each latitudinal zone for all marine data, thus tropical organisms inhabiting deep waters may be subject to low temperatures, potentially increasing their n-3 LC-PUFA content. While we did not specifically account for this in our analysis, we estimated that our dataset of marine fish contained relatively equal proportions of deep-sea fish among the three latitudinal zones (polar 25.6%, temperate 29.2%, tropical 32.5%). Thus, our dataset did not have an undue bias towards "colder-water" deep-sea fish in tropical zones. Further, despite the slightly higher contribution of deep-sea fish in the tropical zone (potentially increasing the n-3 LC-PUFA content in tropical fish), the difference in LC-PUFA content among latitudinal zones is still clearly evident. Thus our analyses yielded a conservative estimate of the observed differences among fishes from different latitudinal zones and especially between polar and tropical fish.

We found that both geographic zone (latitude) and diet (a function of trophic level) can predict FA content in marine organisms. For DHA, carnivores contained the highest levels, which demonstrates selective retention relative to primary producers (Fig. 3). While our results demonstrate and support previous studies that found DHA is selectively retained along the food chain (Kainz et al. 2004; Koussoroplis et al. 2008; Hartwich et al. 2013; Hixson et al. 2015), we also found that FA content was highly dependent on the taxonomic affiliation of organisms, a result which has been previously observed in the taxonomy of phytoplankton (Mourente et al. 1990; Galloway and Winder 2015; Hixson and Arts 2016), zooplankton (Persson and Vrede 2006), and benthic invertebrates (Parrish et al. 1996; Makhutova et al. 2011). It is not well understood if the PUFA content of marine organisms, such as fish, is controlled by phylogenetic, ecological, or trophic factors (Gladyshev et al. 2013); however, according to our analysis of marine organisms in general, these factors are related and interactive.

Fish in polar, temperate, and tropical waters

Fish are of particular interest due to their importance as a protein source and ultimately the main source of pre-formed LC-PUFA for terrestrial vertebrates (including humans). Fish represent a major vector (subsidy) in the transfer of LC-PUFA from marine to terrestrial ecosystems (Gladyshev et al. 2009, 2013). Furthermore, fish and shellfish production in marine ecosystems will likely remain a key source of EPA and DHA for humans. Therefore, it is important to identify and conserve geographic areas that have exceptionally high n-3 LC-PUFA production and subsequent retention in fish. We found that latitudinal geographic zone significantly determined FA content in fish in multivariate space (Fig. 5).

We observed separation of fish in polar and tropical marine zones, with fish in temperate zones in between and dispersed within fish in tropical and polar zones, which created a latitudinal gradient (Fig. 5). Generally, tropical ecosystems contain organisms with FA profiles that are different from those in temperate and polar ecosystems (Iverson 2009), and warm-water fish contain less EPA and DHA and more ARA compared to cold-water fish (Lands 1982; Broadhurst et al. 2002; Kolakowska et al. 2006). The high MUFA levels we observed in polar fish are likely a biomarker of zooplankton (copepods), as this taxonomic group is an important link between lower and higher trophic levels (Parrish 2013). Typical of tropical ecosystems, prey throughout these systems generally contain relatively high levels of n-6 LC-PUFA and very low levels of MUFA (Iverson 2009).

We found that fish in polar and temperate marine zones provide the highest nutritional value for human consumption in terms of n-3 LC-PUFA content. Ironically, some of the most striking impacts of global climate change have been observed in polar oceans (Hoegh-Guldberg and Bruno 2010), and it is known that these regions are particularly sensitive to changes in temperature (Smetacek and Nicol 2005). At the base of marine food webs, the distribution, abundance, and productivity of primary producers (i.e., phytoplankton) are changing in response to eutrophication, warming, acidifying, and stratifying oceans, and the effects of these systemic changes can cascade through food webs (Polovina et al. 2008; Doney et al. 2009). An increase in water temperature due to climate change is predicted to result in decreased proportions of EPA and DHA in a variety of organisms including phytoplankton (Hixson and Arts 2016), copepods (Werbrouck et al. 2016), and albacore tuna (Parrish et al. 2015), as a result of cellular adaptation (homeoviscous adaption) to warming waters. This indicates that the impact of warming waters may alter FA content in consumers, not only through nutrient transfer, but also through their own biochemical response to temperature (Werbrouck et al. 2016).

Compounding these effects, climate change can also shift the normal geographic distribution of fish and other species' populations, as they adapt to current ambient conditions (Stachowicz et al. 2002; Gamito et al. 2015). A rising number of non-polar species are already expanding their ranges into more polar areas (Hoegh-Guldberg and Bruno 2010). Tropical species are predicted to expand toward temperate regions, and temperate species in turn will move into polar regions (Cheung et al. 2009; Vinagre et al. 2011). The dominance of lipid-rich species (with high n-3 LC-PUFA contents) at high latitudes will probably decrease in response to this shift (Gamito et al. 2015). These shifts in population distribution and changes in FA quantity and quality in response to warming temperatures can ultimately affect the health-benefits of commercially relevant species that inhabit the most sensitive regions to climate change.

Implications for consumers

Vertebrates (and most aquatic invertebrates) require DHA (and EPA to some extent) for optimal functioning (Arts et al. 2009; Calder 2015). Endogenous biosynthesis from ALA (or from other precursors, such as EPA or 22:5n-3) is generally limited in most vertebrates, including humans (Brenna et al. 2009). Our data synthesis shows that primary producers in marine and terrestrial ecosystems exert a high level of control over the distribution of EPA and DHA among higher trophic level organisms in their respective food webs. Ultimately EPA and DHA synthesized by aquatic primary producers make their way, by various vectors, to terrestrial organisms (including humans). These n-3 LC-PUFA are not produced in quantity by most plants in terrestrial food webs, therefore preformed EPA and DHA are generally not available to consumers from this source. While terrestrial predators usually cannot obtain EPA and DHA from terrestrial plants, ultimately they can be derived from terrestrial ALA and subsequently bioconverted, either by the predators themselves or by their prey. This indirect route has energetic costs in terms of trophic efficiency, and in terms of the amount of ALA (from plants) and n-3 LC-PUFA derived from their prey because n-3 PUFA are not as prevalent in terrestrial organisms compared to their aquatic counterparts. Thus the supply of preformed EPA and DHA from aquatic organisms generally provides a more abundant and more efficient source of n-3 LC-PUFA to terrestrial animals that have access to aquatic ecosystems (Gladyshev et al. 2009, 2013). Consumers that do not have access to aquatic resources, particularly primary consumers, have undoubtedly faced strong selection pressure to modify ALA and LNA into LC-PUFA, as the ability to synthesize the LC-PUFA is likely related to the inverse of access to these LC-PUFA (Hixson et al. 2015).

Conclusions

We have illustrated a fundamental dichotomy in PUFA abundance and distribution between marine and terrestrial ecosystems. Biome, taxonomy, and trophic level defined FA content of organisms between (and within) ecosystems. Individual n-3 and n-6 PUFA showed different relationships among trophic levels, indicating that EPA, DHA, and ARA are functionally different, and are likely retained or catabolized differently as a function of taxonomy and diet. In marine ecosystems, the n-3 LC-PUFA were higher in polar and temperate organisms than those in the tropics. In particular, fish located in polar and temperate latitudes feature higher levels of n-3 LC-PUFA than those in the tropics. Therefore, we conclude that cool water fish in particular provide an important resource (i.e., essential n-3 LC-PUFA) to many consumers, including terrestrial predators. Climate warming and other anthropogenic influences threaten LC-PUFA production in marine ecosystems in a variety of ways. These factors include increasing water temperatures and its effect on homeoviscous adaptation of cell membranes, biological impacts such as eutrophication (which leads to algal communities being dominated by LC-PUFA-impoverished cyanobacteria), novel species assemblages (i.e., redistribution of species), overfishing (Hoegh-Guldberg and Bruno 2010), and increased absorption of anthropogenic CO2 leading to ocean acidification (Bermúdez et al. 2015). Ultimately these human-induced changes can impact the production, distribution, and transfer of EPA and DHA within the marine ecosystem and may eventually reduce their transfer (subsidy) to terrestrial ecosystems. Therefore, in light of these effects it is becoming increasingly imperative to conserve our marine resources so as to sustain adequate levels of n-3 LC-PUFA for our health and for the health of all animals on this planet.

Acknowledgements

Funding for this work was provided by Ryerson University and by a Natural Sciences and Engineering Research Council Discovery Grant (#04537–2014 to MTA). AW was supported by the German Research Foundation (DFG WA2445/8-1) and MJK by the Austrian Science Fund (FWF project I 1057-B25). The authors thank Alina Tsimbaliouk, Ryerson University, Department of Chemistry and Biology, and Jennifer Weise, Potsdam University, Theoretical Aquatic Ecology and Ecophysiology, for assistance in data collection. Please note that Stefanie M. Colombo was listed in previous publications (from 2010–2015) as Stefanie M. Hixson.

References

- Ackman, R.G., Tocher, C.S., and McLachlan, J. 1968. Marine phytoplankter fatty acids. J. Fish Res. Board Can. 25: 1603–1620. doi:10.1139/f68-145.
- Ahlgren, G., Vrede, T., and Goedkoop, W. 2009. Fatty acid ratios in freshwater fish, zooplankton, and zoobenthos Are there specific optima? *In* Lipids in Aquatic Ecosystems. *Edited by* M.T. Arts, M.T. Brett, and M.J. Kainz. Springer, New York, pp. 147–168.
- Arts, M.T., and Kohler, C.C. 2009. Health and condition in fish: the influence of lipids on membrane competency and immune response. In Lipids in Aquatic Ecosystems. Edited by M.T. Arts, M.T. Brett, and M.J. Kainz. Springer, New York. pp. 237–256.
- Arts, M.T., Ackman, R.G., and Holub, B.J. 2001. 'Essential fatty acids' in aquatic ecosystems: a crucial link between diet and human health and evolution. Can. J. Fish Aquat. Sci. **58**: 122–137. doi:10.1139/f00-224.
- Arts, M.T., Brett, M.T., and Kainz, M.J. 2009. Lipids in aquatic ecosystems. Springer, New York.
- Bazinet, R.P., and Layé, S. 2014. Polyunsaturated fatty acids and their metabolites in brain function and disease. Nat. Rev. Neurosci. 15: 771–785. doi:10.1038/ nrn3820.
- Bermúdez, R., Feng, Y., Roleda, M.Y., Tatters, A.O., Hutchins, D.A., Larsen, T., et al. 2015. Long-Term conditioning to elevated pCO2 and warming influences the fatty and amino acid composition of the diatom Cylindrotheca fusiformis. PLoS ONE, 10(5): e0123945. doi:10.1371/journal.pone.0123945. PMID: 25970340.
- Boelen, P., van Dijk, R., Damsté, J.S.S., Rijpstra, W.I.C., and Burna, A.G.J. 2013. On

- the potential application of polar and temperate marine microalgae for EPA and DHA production. AMB Express, 3: 26. doi:10.1186/2191-0855-3-26. PMID: 23673125
- Bray, J.R., and Curtis, J.T. 1957. An ordination of the upland forest communities of southern Wisconsin. Ecol. Monogr. 27: 325–349. doi:10.2307/1942268.
- Brenna, J.T., and Carlson, S.E. 2014. Docosahexaenoic acid and human brain development: evidence that a dietary supply is needed for optimal development. J. Human Evol. 77: 99–106. doi:10.1016/j.jhevol.2014.02.017.
- Brenna, J.T., Salem, N., Sinclair, A., and Cunanne, S.C. 2009. α-Linolenic acid supplementation and conversion to n-3 long-chain polyunsaturated fatty acids in humans. Prostag. Leukotr. Ess. 80: 85–91. doi:10.1016/j.plefa.2009.01. 004
- Brett, M.T., and Müller-Navarra, D.C. 1997. The role of highly unsaturated fatty acids in aquatic foodweb processes. Freshw. Biol. 38: 483–499. doi:10.1046/j. 1365-2427.1997.00220.x.
- Brett, M.T., Müller-Navarra, D.C., and Persson, J. 2009. Crustacean zooplankton fatty acid composition. *In Lipids in Aquatic Ecosystems*. *Edited by M.T. Arts*, M.T. Brett, and M.J. Kainz. Springer, New York. pp. 115–146.
- Broadhurst, C.L., Wang, Y., Crawford, M.A., Cunnane, S.C., Parkington, J.E., and Schmidt, W.F. 2002. Brain-specific lipids from marine, lacustrine, or terrestrial food resources: potential impact on early African Homo sapiens. Comp. Biochem. Physiol. B: Biochem. Mol. Biol. 131: 653–673. doi:10.1016/S1096-4959(02)00002-7. PMID:11923081.
- Brockerhoff, H., Ackman, R.G., and Hoyle, R.J. 1963. Specific distribution of fatty acids in marine lipids. Arch. Biochem. Biophys. 100: 9–12. doi:10.1016/0003-9861(63)90026-2. PMID:14015630.
- Budge, S.M., Iverson, S.J., Bowen, W.D., and Ackman, R.G. 2002. Among- and within- species variability in fatty acid signatures of marine fish and invertebrates on the Scotian Shelf, Georges Bank, and southern Gulf of St. Lawrence. Can. J. Fish. Aquat. Sci. 59: 886–898. doi:10.1139/f02-062.
- Budge, S.M., Iverson, S.J., and Koopman, H.N. 2006. Studying trophic ecology in marine ecosystems using fatty acids: A primer on analysis and interpretation. Mar. Mam. Sci. 22: 759–801. doi:10.1111/j.1748-7692.2006.00079.x.
- Calder, P.C. 2015. Marine omega-3 fatty acids and inflammatory processes: effects, mechanisms and clinical relevance. Biochim. Biophys. Acta: Mol. Cell Biol. Lipids, 1851: 469–484. doi:10.1016/j.bbalip.2014.08.010. PMID:25149823.
- Cheung, W.W.L., Lam, V.W.Y., Sarmiento, J.L., Kearney, K., Watson, R., and Pauly, D. 2009. Projecting global marine biodiversity impacts under climate change scenarios. Fish Fish. 10: 235–251. doi:10.1111/j.1467-2979.2008.00315.x.
- Clarke, A. 2003. Costs and consequences of evolutionary temperature adaptation. Trends Ecol. Evol. 18: 573–581. doi:10.1016/j.tree.2003.08.007.
- Clarke, K., and Warwick, R. 2001. Change in marine communities: an approach to statistical analyses and interpretation. Primer-E Ltd., Plymouth, UK.
- Cook, H.W., and McMaster, C.R. 2004. Fatty acid desaturation and chain elongation in eukaryotes. *In Biochemistry of lipids*, lipoproteins and membranes. New Comprehensive Biochemistry. *Edited by D.E. Vance and J.E. Vance. Elsevier*. Amsterdam, the Netherlands, pp. 181–204.
- Cossins, A.R., and Prosser, C.L. 1978. Evolutionary adaptation of membrane to temperature. Proc. Natl. Acad. Sci. 75: 2040–2043. doi:10.1073/pnas.75.4.2040. PMID:273929.
- Dalsgaard, J., St. John, M., Kattner, G., Müller-Navarra, D., and Hagen, W. 2003. Fatty acid trophic markers in the pelagic marine food environment. Adv. Mar. Biol. 46: 225–340. doi:10.1016/S0065-2881(03)46005-7. PMID:14601414.
- Domenichiello, A.F., Kitson, A.P., and Bazinet, R.P. 2015. Is docosahexaenoic acid synthesis from α-linolenic acid sufficient to supply the adult brain? Prog. Lipid Res. **59**: 54–66. doi:10.1016/j.plipres.2015.04.002. PMID:25920364.
- Doney, S., Fabry, V., Feely, R., and Kleypas, J. 2009. Ocean acidification: the other CO2 problem. Annu. Rev. Mar. Sci. 1: 169–192. doi:10.1146/annurev.marine. 010908.163834.
- Dubois, V., Breton, S., Linder, M., Fanni, J., and Parmentier, M. 2007. Fatty acid profiles of 80 vegetable oils with regard to their nutritional potential. Eur. J. Lipid. Sci. Technol. **109**: 710–732. doi:10.1002/ejlt.200700040.
- Emery, J.A., Norambuena, F., Trushenski, J., and Turchini, G.M. 2016. Uncoupling EPA and DHA in fish nutrition: Dietary demand is limited in Atlantic salmon and effectively met by DHA alone. Lipids. doi:10.1007/s11745-016-4136-y.
- Galloway, A.W.E., and Winder, M. 2015. Partitioning the relative importance of phylogeny and environmental conditions on phytoplankton fatty acids. PLoS One, 10: e0130053. doi:10.1371/journal.pone.0130053. PMID:26076015.
- Galván, I., Naudí, A., Erritzøe, J., Møller, A.P., Barja, G., and Pamplona, R. 2015. Long lifespans have evolved with long and monounsaturated fatty acids in birds. Evolution, **69**: 2776–2784. doi:10.1111/evo.12754. PMID:26294378.
- Gamito, R., Costa, M.J., and Cabral, H.N. 2015. Fisheries in a warming ocean: trends in fish catches in the large marine ecosystems of the world. Reg. Environ. Change, **15**: 57–65. doi:10.1007/s10113-014-0615-y.
- Gaston, K.J. 2003. The structure and dynamics of geographic ranges. Oxford University Press, New York, NY, U.S.A.
- Gladyshev, M., Arts, M.T., and Sushchik, N. 2009. Preliminary estimates of the export of omega-3 highly unsaturated fatty acids (EPA+DHA) from aquatic to terrestrial ecosystems. In Lipids in Aquatic Ecosystems. *Edited by M.T. Arts, M.T. Brett, and M.J. Kainz. Springer, New York. pp. 179–210.*
- Gladyshev, M.I., Semenchenko, V.P., Dubovskaya, O.P., Fefilova, E.B.,

Makhutova, O.N., Buseva, Z.F., et al. 2011. Effect of temperature on contents of essential highly unsaturated fatty acids in freshwater zooplankton. Limnol. – Ecol. Manage. Inaland Waters, 41: 339–347. doi:10.1016/j.limno.2011.03.001.

- Gladyshev, M.I., Sushchik, N.N., and Makhutova, O.N. 2013. Production of EPA and DHA in aquatic ecosystems and their transfer to the land. Prostaglandins Other Lipid Mediat. **107**: 117–126. doi:10.1016/j.prostaglandins.2013.03.002. PMID:23500063.
- Glencross, B.D., De Santis, C.D., Bicskei, B., Taggart, J.B., Bron, J.E., Betancor, M.B., and Tocher, D.R. 2015. A comparative analysis of the response of the hepatic transcriptome to dietary docosahexaenoic acid in Atlantic salmon (*Salmo salar*) post-smolts. BMC Genom. **16**: 684. doi:10.1186/s12864-015-1810-z.
- Greenacre, M., and Primicerio, R. 2013. Multivariate analysis of ecological data. Fundación BBVA. Bilbao Spain. pp. 279–281.
- Guglielmo, C.G. 2010. Move that fatty acid: Fuel selection and transport in migratory birds and bats. Integr. Comp. Biol. **50**: 336–345. doi:10.1093/icb/icq097. PMID:21558208.
- Guschina, I.A., and Harwood, J.L. 2006. Mechanisms of temperature adaptation in poikilotherms. FEBS Lett. **580**: 5477–5483. doi:10.1016/j.febslet.2006.06. 066. PMID:16824520.
- Hartwich, M., Martin-Creuzburg, D., and Wacker, A. 2013. Seasonal changes in the accumulation of polyunsaturated fatty acids in zooplankton. J. Plankton Res. 35(1): 121–134. doi:10.1093/plankt/fbs078.
- Hazel, J.R. 1995. Thermal adaptation in biological membranes: is homeoviscous adaptationthe explanation? Annu. Rev. Physiol. 57:19–42. doi:10.1146/annurev. ph.57.030195.000315. PMID:7778864.
- Hixson, S.M., and Arts, M.T. 2016. Climate warming is predicted to reduce omega-3, long-chain, polyunsaturated fatty acid production in phytoplankton. Global Change Biol. 22: 8. doi:10.1111/gcb.13295.
- Hixson, S.M., Sharma, B., Kainz, M.J., Wacker, A., and Arts, M.T. 2015. Production, distribution, and abundance of long-chain omega-3 polyunsaturated fatty acids: A fundamental dichotomy between freshwater and terrestrial ecosystems. Environ. Rev. 23: 414–424. doi:10.1139/er-2015-0029.
- Hoegh-Guldberg, O., and Bruno, J.F. 2010. The impact of climate change on the world's marine ecosystems. Science, 5985: 1523–1528. doi:10.1126/science. 1189930.
- Hong, S., Gronert, K., Devchand, P.R., Moussignac, R.L., and Serhan, C.N. 2003. Novel docosatrienes and 17S-resolvins generated from docosahexaenoic acid in murine brain, human blood, and glial cells. Autacoids in antiinflammation. J. Biol. Chem. 278: 14677–14687. doi:10.1074/jbc.M300218200.
- Iverson, S.J. 2009. Tracing aquatic food webs using fatty acids: from qualitative indicators to quantitative determination. *In Lipids in Aquatic Ecosystems*. *Edited by M.T. Arts, M.T. Brett, and M.J. Kainz. Springer, New York. pp. 281–307.*
- Iverson, S.J., Frost, K.J., and Lowry, L.F. 1997. Fatty acid signatures reveal fine scale structure of foraging distribution of harbor seals and their prey in Prince William Sound, Alaska. Mar. Ecol. Prog. Ser. 151: 255–271. doi:10.3354/ meps151255.
- Iverson, S.J., Field, C., Bowen, W.D., and Blanchard, W. 2004. Quantitative fatty acid signature analysis: a new method of estimating predator diets. Ecol. Monogr. 74: 211–235. doi:10.1890/02-4105.
- Kainz, M.J., Arts, M.T., and Mazumber, A. 2004. Essential fatty acids in the planktonic food web and their ecological role for higher trophic levels. Limnol. Oceanogr. 49: 1784–1793. doi:10.4319/lo.2004.49.5.1784.
- Kattner, G., and Hagen, W. 2009. Lipids in marine copepods: Latitudinal characteristics and perspective to global warming. In Lipids in Aquatic Ecosystems. Edited by M.T. Arts, M.T. Brett, and M.J. Kainz. Springer, New York. pp. 257–280.
- Kelly, J.R., and Scheibling, R.E. 2012. Fatty acids as dietary tracers in benthic food webs. Mar. Ecol. Prog. Ser. 446: 1–22. doi:10.3354/meps09559.
- Kolakowska, A., Domiszewski, Z., and Bienkiewicz, G. 2006. Effects of biological and technological factors on the utility of fish as a source of n-3 PUFA. In Omega-3 Fatty Acid Research. Edited by M.C. Teale. Nova Science Publishers Inc. New York. pp. 83–107.
- Koussoroplis, A., Lemarchand, C., Bec, A., Desvilettes, C., Amblard, C., Fournier, C., et al. 2008. From aquatic to terrestrial food webs: decrease of the docosahexaenoic acid/linoleic acid ratio. Lipids, 43: 461–466. doi:10.1007/s11745-008-3166-5. PMID:18335265.
- Lands, W.E.M. 1982. Biochemical observations on long-chain fatty acids from fish oil and their effect on prostaglandin synthesis in animals and humans. *In* Nutritional Evaluation of Long Chain Fatty Acids in Fish Oil. *Edited by* S.M. Barlow and M.E. Stansby. Academic Press, London. pp. 25–88.
- Lane, R.L., Trushenski, J.T., and Kohler, C.C. 2006. Modification of fillet composition and evidence of differential fatty acid turnover in sunshine bass Morone chrysops × M. saxatilis following change in dietary lipid source. Lipids, 41: 1029–1038. PMID:17263302.
- Maillet, D., and Weber, J.M. 2006. Performance-enhancing role of dietary fatty acids in a long distance migrant shorebird: the semi palmated sandpiper. J. Exp. Biol. 209: 2686–2695. doi:10.1242/jeb.02299. PMID:16809459.
- Makhutova, O.N., Sushchik, N.N., Gladyshev, M.I., Ageev, A.V., Pryanichnikova, E.G., and Kalachova, G.S. 2011. Is the fatty acid composition of freshwater zoobenthic invertebrates controlled by phylogenetic or trophic factors? Lipids, 46: 709–721. doi:10.1007/s11745-011-3566-9. PMID:21567302.

Mourente, G., Lubián, L.M., and Odriozola, J.M. 1990. Total fatty acid composition as a taxonomic index of some marine microalgae used as food in marine aquaculture. Hydrobiologia, 203: 147–154. doi:10.1007/BF00005683.

- Murray, D.S., Hager, H.H., Tocher, D.R., and Kainz, M.J. 2014. Effect of partial replacement of dietary fish meal and oil by pumpkin kernel cake and rape-seed oil on fatty acid composition and metabolism in Arctic charr (*Salvelinus alpinus*). Aquaculture, 431: 85–91. doi:10.1016/j.aquaculture.2014.03.039.
- Parrish, C.C. 2009. Essential fatty acids in aquatic food webs. In Lipids in Aquatic Ecosystems. Edited by M.T. Arts, M.T. Brett, and M.J. Kainz. Springer. New York, pp. 309–326.
- Parrish, C.C. 2013. Lipids in marine ecosystems. [Online]. ISRN Oceanography. doi:10.5402/2013/604045.
- Parrish, C.C., Yang, Z., Lau, A., and Thompson, R.J. 1996. Lipid composition of Yoldia hyperborea (Protobranchia), Nephthys ciliata (Nephthydae) and Artacama proboscidea (Terebellidae) living at sub-zero temperatures. Comp. Biochem. Physiol. B: Biochem. Mol. Biol. 114: 59–67. doi:10.1016/0305-0491(95)02117-5.
- Parrish, C.C., Pethybridge, H., Young, J.W., and Nichols, P.D. 2015. Spatial variation in fatty acid trophic markers in albacore tuna from the southwestern Pacific Ocean A potential 'tropicalization' signal. Deep Sea Res. II: Topical Stud. Oceanogr. 113: 199–207. doi:10.1016/j.dsr2.2013.12.003.
- Persson, J., and Vrede, T. 2006. Polyunsaturated fatty acids in zooplankton: variation due to taxonomy and trophic position. Freshw. Biol. **51**: 887–900. doi:10.1111/j.1365-2427.2006.01540.x.
- Polovina, J.J., Howell, E.A., and Abecassis, M. 2008. Ocean's least productive waters are expanding. Geophys. Res. Lett. 35: L03618. doi:10.1029/2007GL031745.
- Rombenso, A.N., Trushenski, J.T., Jirsa, D., and Drawbridge, M. 2016. Docosahexaenoic acid (DHA) and arachidonic acid (ARA) are essential to meet LC-PUFA requirements of juvenile California Yellowtail (Seriola dorsalis). Aquaculture, 463: 123–134. doi:10.1016/j.aquaculture.2016.05.004.
- Schlechtriem, C., Arts, M.T., and Zellmer, I.D. 2006. Effect of temperature on the fatty acid composition and temporal trajectories of fatty acids in fasting Daphnia pulex (Crustacea, Cladocera). Lipids, 41: 397–400. doi:10.1007/s11745-006-5111-9. PMID:16808154.
- Simopoulos, A.P. 2011. Evolutionary aspects of diet: the omega-6/omega-3 ratio and the brain. Mol. Neurobiol. **44**: 203–215. doi:10.1007/s12035-010-8162-0. PMID:21279554
- Sinensky, M. 1974. Homeoviscous adaptation Homeostatic process that regulates viscosity of membrane lipids in *Escherichia coli*. Proc. Natl. Acad. Sci. **71**: 522–525. doi:10.1073/pnas.71.2.522. PMID:4360948.
- Skjervold, H. 1992. Lifestyle diseases and the human diet. How should the new discoveries influence the food production? *In* The Journal of Dairy Industry of Norway. Norway, p. 48.
- Smetacek, V., and Nicol, S. 2005. Polar ocean ecosystems in a changing world. Nature, 437: 362–368. doi:10.1038/nature04161.
- Sperfeld, E., and Wacker, A. 2012. Temperature affects the limitation of *Daphnia magna* by eicosapentaenoic acid, and the fatty acid composition of body tissue and eggs. Freshw. Biol. **57**: 497–508. doi:10.1111/j.1365-2427.2011.02719.x.
- Stachowicz, J.J., Terwin, J.R., Whitlatch, R.B., and Osman, R.W. 2002. Linking climate change and biological invasions: Ocean warming facilitates nonindigenous species invasions. Proc. Natl. Acad. Sci. 99: 15497–15500. PMID: 12422019.
- Swanson, D., Block, R., and Mousa, S.A. 2012. Omega-3 fatty acids EPA and DHA: Health benefits throughout life. Adv. Nutr. Int. Rev. J. 3: 1–7. doi:10.3945/an. 111.000893. PMID:22332096.
- Taipale, S., Strandberg, U., Peltomaa, E., Galloway, A.W.E., Ojala, A., and Brett, M. 2013. Fatty acid composition as biomarkers of freshwater microalgae: analysis of 37 strains of microalgae in 22 genera and in seven classes. Aquat. Microb. Ecol. 71: 165–178. doi:10.3354/ame01671.
- Thomas, D.N., and Dieckmann, G.S. 2002. Antarctic sea ice a habitat for extremophiles. Science, 295: 641–644. doi:10.1126/science.1063391. PMID:11809961.
- Tocher, D.R., Dick, J.R., MacGlaughlin, P., and Bell, J.G. 2006. Effect of diets enriched in Delta 6 desaturated fatty acids (18:3n-6 and 18:4n-3), on growth, fatty acid composition and highly unsaturated fatty acid synthesis in two populations of Arctic charr (Salvelinus alpinus L.). Comp. Biochem. Physiol. B: Biochem. Mol. Biol. 144: 245–253. doi:10.1016/j.cbpb.2006.03.001. PMID:16630735.
- Torres-Ruiz, M., Wehr, J.D., and Perrone, A.A. 2010. Are net-spinning caddisflies what they eat? An investigation using controlled diets and fatty acids. J. N. Am. Benthol. Soc. 29: 803–813. doi:10.1899/09-162.1.
- Trushenski, J., Schwarz, M., Bergman, A., Rombenso, A., and Delbos, B. 2012. DHA is essential, EPA appears largely expendable, in meeting the n-3 long-chain polyunsaturated fatty acid requirements of juvenile cobia Rachycentron canadum. Aquaculture, 326–329: 81–89. doi:10.1016/j.aquaculture.2011.11.033.
- Turcotte, C., Chouinard, F., Lefebvre, J.S., and Flamand, N. 2015. Regulation of inflammation by cannabinoids, the endocannabinoids 2-arachidonylglycerol and arachidonoyl-athanolamide, and their metabolites. J. Leukoc. Biol. 97: 1049–1070. doi:10.1189/jlb.3RU0115-021R. PMID:25877930.
- Twining, C.W., Brenna, J.T., Hairston, N.G., and Flecker, A.S. 2016. Highly unsaturated fatty acids in nature: what we know and what we need to learn. Oikos, 125: 749–760. doi:10.1111/oik.02910.
- van Ginneken, V.J.T., Helsper, J.P.F.G., de Visser, W., van Keulen, H., and

Brandenburg, W.A. 2011. Polyunsaturated fatty acids in various macroalgal species from north Atlantic and tropical seas. Lipids Health Dis. **10**: 104. doi:10.1186/1476-511X-10-104. PMID:21696609.

Vinagre, C., Santos, F.D., Cabral, H., and Costa, M.J. 2011. Impact of climate warming upon the fish assemblages of the Portuguese coast under different scenarios. Reg. Environ. Change, 11: 779–789. doi:10.1007/s10113-011-0215-z.

Wacker, A., and Martin-Creuzburg, D. 2007. Allocation of essential lipids in

Daphnia magna during exposure to poor food quality. Funct. Ecol. **21**: 738–747. doi:10.1111/j.1365-2435.2007.01274.x.

Werbrouck, E., Van Gansbeke, D., Vanreusel, A., Mensens, C., and De Troch, M. 2016. Temperature-induced changes in fatty acid dynamics of the intertidal grazer *Platychelipus littoralis* (Crustacea, Copepoda, Harpacticoida): insights from a short-term feeding experiment. J. Thermal Biol. **57**: 44–53. doi:10.1016/j.jtherbio.2016.02.002.